



Article

Temporal Variation of NO₂ and HCHO Vertical Profiles Derived from MAX-DOAS Observation in Summer at a Rural Site of the North China Plain and Ozone Production in Relation to HCHO/NO₂ Ratio

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Abstract: We performed a comprehensive and intensive field experiment including ground-based multi-axis differential optical absorption spectroscopy (MAX-DOAS) measurement at Raoyang (115°44′ E, 38°14′ N; 20 m altitude) in summer (13 June–20 August) 2014. The NO₂ and HCHO profiles retrieved by MAX-DOAS take on different vertical distribution shapes, with the former declining with the increasing altitude and the latter having an elevated layer. The average levels of vertical column densities (VCDs) and near-surface volume mixing ratios (VMRs) were $1.02\pm0.51 imes10^{16}$ molec·cm $^{-2}$ and 3.23 \pm 2.70 ppb for NO₂ and 2.32 \pm 0.56 \times 10¹⁶ molec·cm⁻² and 5.62 \pm 2.11 ppb for HCHO, respectively. The NO2 and HCHO levels are closely connected with meteorological conditions, with the larger NO₂ VCDs being associated with lower temperature, higher relative humidity (RH) and lower planetary boundary layer height (PBLH). With respect to the diurnal variations of vertical distribution, the NO₂ in the residual layer gradually disappeared from 1.2 km height to the surface during the period of 7:00-11:00 Beijing time (BJ), and the near-surface NO2 had larger VMRs in the early morning and evening than in the later morning and afternoon. An elevated HCHO layer was observed to occur persistently with the lifted layer height rising from ~0.5 km to ~1.0 km before 10:00 BJ; the near-surface HCHO VMRs gradually increased and peaked around 10:00 BJ. The ratios of HCHO to NO₂ (R_{HCHO-NO2}) were generally larger than two in the boundary layer from 11:00 BJ until 19:00 BJ, the time period when ozone photochemistry was most active. Thus, ozone (O₃) production was mainly in the NOx-limited regime during the observation campaign, which was closely related to relatively high temperatures and low RH. The O3 production regimes also changed with the wind's direction. These results are significant to reveal the formation mechanism of O₃ pollution and develop strategies for controlling the O₃ photochemical pollution over the North China Plain.

Keywords: NO₂; HCHO; vertical profile; ozone production; MAX-DOAS



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1. Introduction

Nitrogen dioxide (NO_2) and formaldehyde (HCHO) are not only important air pollutants in the troposphere, but also crucial precursors of ozone which play a key role in atmospheric chemistry [1–5]. Atmospheric NO_2 , a kind of nitrogen oxide (NO_x), is generated by human activity (mainly fossil fuel combustion and biomass combustion) and natural processes (such as microbial processes and lightning) [6–8]. HCHO comes not only from primary sources, such as the emissions of biomass combustion, transportation and industry, but also secondary sources, i.e., photochemical production, by the way of the oxidation of volatile organic compounds (VOCs) [9]. Therefore, NO_2 and HCHO are

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usually used as the markers of NO_x and VOCs, respectively [10]. The investigations of NO_2 and HCHO vertical distribution as well as their ratios are beneficial to reveal the spatial-temporal evolution of ozone- NO_x -VOC sensitivity and formulate a reasonable ozone (O₃) pollution control strategy [11,12].

The North China Plain (NCP) is one of the high-risk areas of O₃ pollution, and O₃ spatiotemporal heterogeneity is impacted by complex factors such as the photochemical process [13]. The limited vertical observations of NO₂ and HCHO restrict the comprehensive understanding of the O₃ pollution mechanism and long-term trend [14]. Ground-based multi-axis differential optical absorption spectroscopy (MAX-DOAS) has been widely used as a stereoscopic monitoring method to obtain NO₂ and HCHO profiles [15]. For example, the MAX-DOAS observation at a rural site in the central-western NCP found lifted layers for both NO₂ and HCHO with the 20–30% contribution of regional transport and the influence of agricultural burning on HCHO variation in May and June 2016 [16]. The ratios of HCHO to NO₂ concentrations retrieved from (MAX-)DOAS measurement are usually applied to analyze the ozone-NO_x-VOC sensitivity in different regions of China [17–21]. This research has shown that there are significant spatial and temporal variations for O₃ photochemical production regimes. In addition, the studies by the observation-based box model suggest that in situ O₃ photochemical formation varies in different controlled regimes at two rural NCP sites due to the polluted plume transport and biogenic emissions [22,23]. Moreover, the research at another NCP site indicates the regimes of O₃ production changing with height and pollution conditions because of different levels of O_3 precursors [24]. Despite the above studies, our knowledge of the vertical distribution of temporal evolutions of NO2 and HCHO and their influence on O₃ production is still insufficient for us to understand the formation mechanism of O₃ pollution and develop strategies for controlling the O₃ pollution in the NCP rural area.

In this study, we present NO_2 and HCHO vertical distributions retrieved from ground-based MAX-DOAS observation during a campaign in summer 2014 at a rural site of the NCP. Section 2 introduces the site, measurement instruments, methods of spectral analysis and vertical profile retrieval of NO_2 and HCHO, as well as ancillary datasets. Section 3 shows the overall characteristics and diurnal variations of vertical profiles of NO_2 , HCHO and HCHO/ NO_2 ratio, as well as the relationship between HCHO/ NO_2 ratio and O_3 production. Finally, discussion and conclusions are given in Sections 4 and 5, respectively.

2. Experiments and Methods

2.1. Site and Instrument

As an NCP rural site, Raoyang meteorological station (115°44′ E, 38°14′ N; 20 m above sea level) is located in the middle of Hebei Province, China (Figure 1a). There are no large local industrial emission sources. Agriculture is the main industry in Raoyang County. The Yanshan and Taihang Mountains are distributed in the north and west of Raoyang, respectively. A cluster of industrial and populated cites surrounding Raoyang observatory are approximately 50–200 km away (Figure 1b) [25]. More details of geographical conditions and atmospheric environment concerning Raoyang station can be found in previous studies [26–29].

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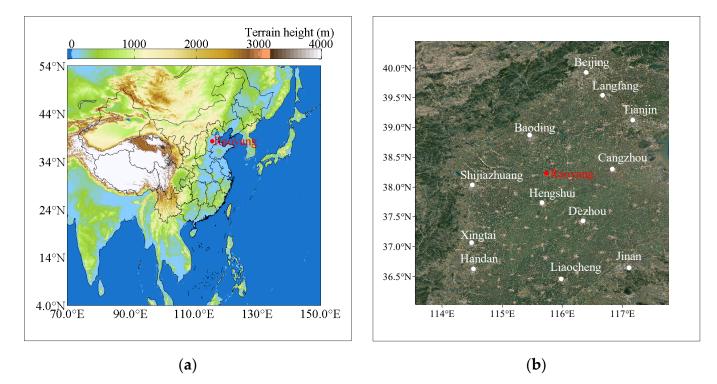


Figure 1. (a) Geographical location of the Raoyang meteorological station in China and (b) satellite image of the major neighborhood cities around Raoyang observatory, marked by white and red dots, respectively.

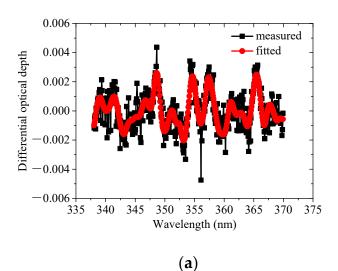
The Mini MAX-DOAS system was set up on the roof of three-story building at Raoyang. It operated automatically from 13 June to 20 August 2014. The compact instrument, made by Hoffmann Messtechnik GmbH in Germany, collected scattered light via a telescope which was coupled to a fiber and transferred to a spectrograph. The spectrograph operated at the temperature of 5 °C, covering the wavelength range of 292–447 nm. A stepper motor drove the instrument to achieve the measurement at different elevation angles with the same southeastern azimuth. The instrument recorded the scattered solar spectra at 11 elevation angles $(1^{\circ}, 2^{\circ}, 3^{\circ}, 4^{\circ}, 5^{\circ}, 6^{\circ}, 8^{\circ}, 10^{\circ}, 15^{\circ}, 30^{\circ}$ and 90°). The integration time of each individual spectrum was \sim 0.5 min. We performed spectral correction with the spectra of dark current and electronic offset as well as wavelength calibration before spectral fitting. A computer with professional software was used for operational control and data collection. More descriptions about the Mini MAX-DOAS instrument are available in our previous work [25,30–33].

2.2. Spectral Analysis

According to the Beer–Lambert law, the differential slant column densities (dSCDs) of target species (such as NO_2 and HCHO in this article) can be retrieved from measured spectra of scattering sunlight by the DOAS method [34]. The NO_2 and HCHO dSCDs denote the differences of their slant column densities (SCDs) between measurement spectra and reference spectra. In this study, reference spectra were derived from sequential spectra, which were defined as the interpolated spectra between two zenith spectra measured before and after an off-zenith sequence of elevation angles. In this case, the NO_2 and HCHO dSCDs from spectral inversion can be treated as tropospheric NO_2 and HCHO dSCDs, i.e., so-called NO_2 and HCHO delta SCDs [35]. The process of spectral analysis was achieved by QDOAS software (http://uv-vis.aeronomie.be/software/QDOAS/) based on the theory of non-linear least squares. The settings of fitting parameters for NO_2 and HCHO spectral analysis can be taken from previous studies [16,31], listed in Table S1. In the post-processing of data quality control, NO_2 and HCHO dSCDs were rejected when the

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root mean square (RMS) of spectral fitting residuals was larger than 0.003 [25]. Figure 2 shows an example of spectral fitting for NO_2 and HCHO dSCDs from spectrum measured at the elevation angle of 10° at 12:05 Beijing time (BJ, UTC + 8 h) on 13 June 2014.



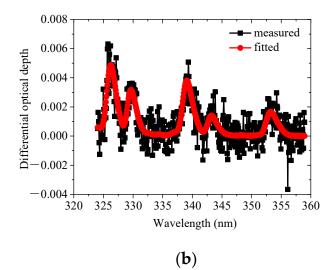


Figure 2. Example of spectral fitting for (a) NO₂ and (b) HCHO. Black and red curves with symbols indicate the measured and fitted differential optical depth, respectively. The fitted NO₂ and HCHO dSCDs are 3.73×10^{16} molec·cm⁻² and 7.07×10^{16} molec·cm⁻², respectively. The corresponding root mean square (RMS) of fitting residuals between measured and fitted spectra are 7.98×10^{-4} and 8.81×10^{-4} , respectively.

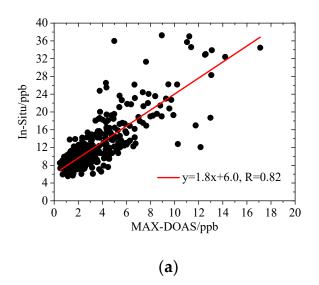
2.3. Retrieval of NO₂ and HCHO Vertical Profiles

Vertical profiles of NO₂ and HCHO volume mixing ratios (VMRs) in the lower troposphere (0-4 km) were separately retrieved from each elevation sequence of NO₂ and HCHO delta dSCDs by the "Profile inversion algorithm of aerosol extinction and trace gas concentration" (PriAM) algorithm, which was a type of optimal estimation (OE) profile inversion algorithm for ground-based MAX-DOAS observation [36]. The solution of PriAM profile retrieval is achieved by numerical iteration procedure. PriAM contains a two-step inversion procedure, i.e., profile inversion of aerosol extinction (AE) and profile inversion of trace gas VMRs. The AE vertical profiles at 360 nm have been successfully retrieved from the oxygen dimer (O_4) delta dSCDs during the field campaign. The specific results and inversion parameters (for example, surface albedo, single scattering albedo and asymmetry factor of aerosol, temperature and pressure profiles) for AE profiles can be referred to previous studies [25]. The AE profiles were used as the input parameters for NO₂ and HCHO profile inversion in the second step. On the basis of the spectral fitting intervals of NO₂ and HCHO, the simulations of air mass factor through the SCIATRAN radiative transfer model (RTM) were carried out at 360 nm for NO₂ and 343 nm for HCHO. Retrieved AE profiles were converted into those at NO₂ and HCHO simulation wavelengths with an Angström exponent of 0.9, estimated from the collocated nephelometer measurement [26]. An exponential a priori NO₂ profile and a Boltzmann distribution a priori HCHO profile the same as that used in the adjacent site were used in this study [16]. The diagonal elements of the a priori covariance matrix were set as 1 for NO₂ and 0.25 for HCHO, which did not decrease with the altitudes in order to balance the flexibility and stability of the profile inversion. Except the first layer (0–50 m) adjacent to the ground, hereafter called the near-surface layer, the vertical interval in the NO₂ and HCHO profiles was set as 200 m. During the procedure of data post processing, we rejected the NO₂ (HCHO) profiles with a cost function of profile inversion larger than 30 and the relative deviations of NO₂ (HCHO) dSCDs between PriAM simulation and MAX-DOAS measurement greater than 30%. The screening thresholds were based on the balance of data quality and amount. With respect to selected thresholds, Atmosphere **2022**, 13, 860 5 of 15

82% NO₂ profiles and 72% HCHO profiles were left, and the corresponding averages of OE cost function were 5.72 for NO₂ and 9.90 for HCHO. Then, the near-surface VMRs and vertical column densities (VCDs) of NO₂ and HCHO could be obtained by extraction and vertical integration through NO₂ and HCHO profiles. In addition, it should be noted that we use all the qualified MAX-DOAS data below without the classification of sky condition because NO₂ and HCHO results are less influenced under most cloudy conditions (except fog and optically thick clouds) [36]. Meanwhile, the aforementioned data quality procedure can eliminate the parts of data interfered with by different sky conditions.

2.4. Ancillary Datasets

Hourly meteorological data, such as surface temperature, wind and relative humidity, were obtained from the synchronous operational observation of China Meteorological Administration (CMA) at the same site. The ERA5 reanalysis data (https://www.ecmwf.int/en/forecasts/datasets/reanalysis-datasets/era5, accessed on 15 June 2020) were also used for the extraction of the planetary boundary layer height (PBLH) as well as the a priori profiles of temperature and pressure. The measurements of in situ surface O₃, NO₂ and HCHO were also carried out during the intensive field campaign in the summer of 2014 [27,29]. Figure 3 shows the correlation of hourly concentrations between two data sets, i.e., in situ measurements and MAX-DOAS retrieval. The correlation coefficients (R) are 0.82 and 0.46 for NO₂ and HCHO, respectively, which are similar to the results at a station on the central-western edge of the same NCP area [16]. To some extent, this implies that the retrieved data in this study are reliable to investigate the variations in NO₂ and HCHO.



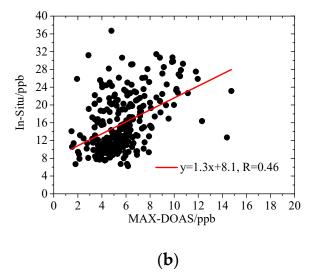


Figure 3. Correlation of the hourly averaged (a) NO₂ and (b) HCHO VMRs (black dots) between in situ measurement and MAX-DOAS retrieval in the near-surface layer. The red lines indicate the linear fit between both data sets. The unit conversion between "ppb" and "molec·cm⁻³" for MAX-DOAS observation is achieved by the coefficient of 2.5×10^{10} .

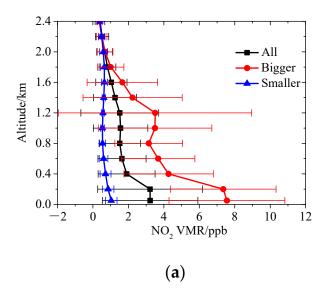
3. Results

3.1. Abundance and Overall Characteristics

Owing to the fact that the substantial information on the concentrations derived from the measurements was mainly below 2 km (Figure S1), the averaged NO₂ and HCHO vertical profiles below 2.4 km during the effective observation period are shown by the black lines with squares in Figure 4. The corresponding means \pm standard deviations of VCDs and near-surface VMRs are $1.02\pm0.51\times10^{16}$ molec·cm $^{-2}$ and 3.23 ± 2.70 ppb for NO₂ and $2.32\pm0.56\times10^{16}$ molec·cm $^{-2}$ and 5.62 ± 2.11 ppb for HCHO. The correlation coefficients between VCDs and near-surface VMRs are 0.69 for NO₂ and 0.55 for HCHO,

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implying that there are limitations in the representation of NO_2 and HCHO variations at higher altitudes due to their surface concentrations. This can be further confirmed by the vertical structure of NO_2 and HCHO. While the NO_2 VMR decreases with the increasing altitude as a whole, there is generally a significantly elevated HCHO pollution layer at the altitude of ~1 km mainly caused by HCHO secondary production at higher altitudes.



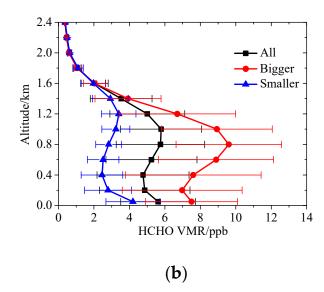


Figure 4. Vertical profiles of (a) NO_2 and (b) HCHO VMRs. The black lines with squares represent the profiles averaged by all the valid NO_2 (HCHO) profiles during the effective observation period. According to the mean and standard deviation of all the valid NO_2 (HCHO) VCDs, the corresponding profiles are averaged on the condition of NO_2 (HCHO) VCDs larger than "mean + standard deviation" (the red lines with dots) or smaller than "mean – standard deviation" (the blue lines with triangles), respectively. The error bars denote the standard deviations of NO_2 (HCHO) VMRs at different altitudes.

According to the aforementioned means and standard deviations of NO₂ (HCHO) VCDs, two classifications of NO₂ (HCHO) profile corresponding to the NO₂ (HCHO) VCD "larger than mean + standard deviation" or "smaller than mean — standard deviation", are screened and averaged, respectively, and presented as "Larger" and "Smaller" in Figure 4. Then, the NO₂ (HCHO) levels and meteorological conditions for the three situations of "All", "Larger", and "Smaller" are also summarized in Table 1. For NO₂, the secondary peak at the higher altitude (~1.1 km) appears to gradually become clear with the increasing NO₂ VCD (Figure 4a). For HCHO, although the peak at the higher altitude (0.8–1.2 km) always exists, it enhances gradually with the increasing HCHO VCD (Figure 4b). The relatively lower temperature and higher relative humidity (RH) favor more severe NO₂ pollution, while there are no significant differences in temperature and RH for the three HCHO situations (Table 1). This is probably connected with the differences of chemical reaction affecting NO2 and HCHO. Through comparing the wind for the "Larger" and "Smaller" situations, we find that the higher VCDs correspond to the lower wind speeds (WS) with the wind directions (WD) of north for NO₂ and southeast for HCHO. In addition, the NO₂ and HCHO pollution levels are closely related to the PBLHs. The higher PBLH, the weaker the NO₂ VCDs.

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Table 1. Averages for the NO₂ and HCHO concentrations as well as corresponding meteorological parameters.

Parameters	NO ₂			НСНО		
	All	Larger	Smaller	All	Larger	Smaller
VCD/10 ¹⁶ molec·cm ⁻²	1.02 ± 0.51	2.03 ± 0.60	0.47 ± 0.04	2.32 ± 0.56	3.29 ± 0.46	1.51 ± 0.18
VMR/ppb	3.23 ± 2.70	7.56 ± 3.27	1.03 ± 0.33	5.62 ± 2.11	7.50 ± 2.59	4.20 ± 1.53
Temperature/°C 30		26	33	30	29	28
RH/%	55	66	42	55	61	58
WD/°	120	353	98	120	131	263
$WS/m \cdot s^{-1}$	2.2	1.8	2.2	2.2	1.6	2.8
PBLH/km	1.16	0.46	1.56	1.16	0.99	0.95

3.2. Diurnal Variations

There are significant differences in the vertical distribution of NO_2 and HCHO diurnal pattern (Figure 5a,b). It is clear that there are elevated pollution layers for both NO_2 and HCHO before 10:00 BJ above the planetary boundary layer, i.e., in the residual layer (Figure 5a,b,h). With the development of a daytime planetary boundary layer, the NO_2 lifted layer gradually disappears, and NO_2 concentrates mainly at the heights of less than 0.2 km from 11:00–19:00 BJ. However, the elevated HCHO layer always exists through the entire day, and the associated HCHO concentration levels are lower due to the higher PBLH from 11:00–17:00 BJ. The most likely cause of elevated HCHO layers during 11:00–17:00 BJ is the more active photochemical reactions occurring in the upper part of the boundary layer, consistent with previous research results on the atmospheric pollution and oxidation pool over NCP [2].

The diurnal variations of NO₂ VCD and near-surface VMR are basically consistent, and there are significant differences in the HCHO diurnal pattern between VCDs and near-surface VMRs (Figure 5c–f). For example, different from the peaks of NO₂ VCDs and near-surface VMRs in the early morning and evening, the HCHO near-surface VMR gradually increases and peaks around 10:00 BJ. The diurnal variations of VCD and near-surface VMR are in the range of 0.74– 1.78×10^{16} molec·cm⁻² and 1.74–7.63 ppb for NO₂ and 2.11– 2.95×10^{16} molec·cm⁻² and 4.67–7.47 ppb for HCHO. Comparing the diurnal variations of temperature and RH with the near-surface VMRs of NO₂ and HCHO (Figure 5e–g), we still find that higher NO₂ levels are associated with lower temperature and higher RH, and there is no significant relationship between HCHO concentration and temperature (or RH). Considering the relatively weak wind conditions (Figure 5h), the peaks of NO₂ VCDs and near-surface VMRs in the early morning and evening are probably caused by local sources such as traffic emission. Meanwhile, the NO₂ and HCHO pollution is relatively weaker when the planetary boundary layer is fully developed around 15:00 BJ, showing a PBLH of ~1.5 km (Figure 5c–f,h).

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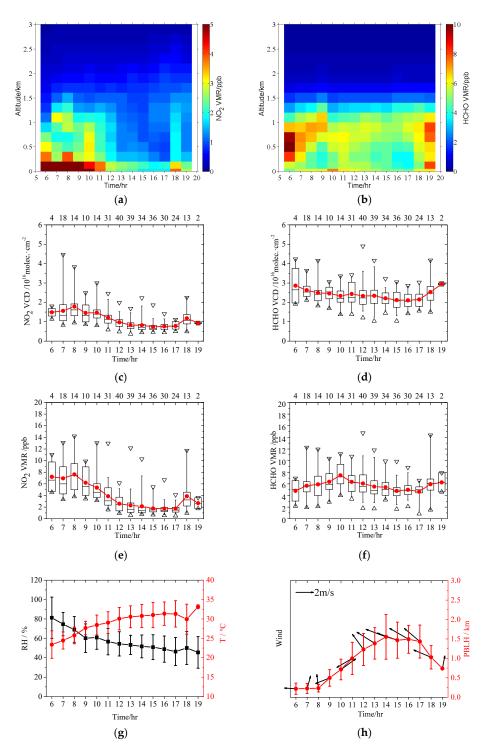


Figure 5. Diurnal variations of (a) NO₂ VMR profiles, (c) NO₂ VCD and (e) NO₂ VMR in the lowest layer based on the qualified hourly MAX-DOAS data as well as coincident meteorological conditions; (g) diurnal pattern of in situ surface RH and T; (h) diurnal variations of in situ wind direction and speed presented by vector arrow and PBLH from linearly interpolated reanalysis data. (b,d,f) are separately the same as (a,c,e) except for HCHO. Lower (upper) error bars, boxes and lower (upper) triangles in (c-f) are the 5th (95th), 25th (75th) percentiles, minima (maxima) of the data grouped in each hour, respectively. Hyphens inside the boxes and red curves with circles in (c-f) separately denote the medians and the averages. The numbers of integrated sampling days for specific hour are labeled at the top axis in (c-f). The error bar in (g,h) denotes the standard deviation, and the scale of wind vector is also shown by a horizontal arrow in (h). The time shown in the figures is Beijing time (BJ).

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3.3. HCHO/NO₂ Ratio

3.3.1. Vertical Profile and Diurnal Variation

The ratio of HCHO to NO₂ ($R_{HCHO-NO2}$) has been proved to be an effective indicator to investigate the sensitivity of tropospheric ozone production to nitrogen oxide and reactive volatile organic compound (VOC) [37,38]. The averaged $R_{HCHO-NO2}$ profile (Figure 6a) was calculated through the profiles of HCHO and NO₂ VMRs retrieved from ground-based MAX-DOAS. For the calculation of $R_{HCHO-NO2}$ (Figure S2), only the data with HCHO VMRs larger than 1 ppb and smaller than 10 ppb were considered in order to balance the data reliability and amount of $R_{HCHO-NO2}$. The averages of $R_{HCHO-NO2}$ calculated from near-surface VMRs and VCDs are $R_{HCHO-NO2,VMR} = 2.71 \pm 1.67$ and $R_{HCHO-NO2,VCD} = 2.61 \pm 0.94$, respectively, during the effective observation period; these values are close to each other. However, the $R_{HCHO-NO2}$ below 2.0 km is in the range of 0.88–5.39 with an elevated layer around the altitude of 1.0 km.

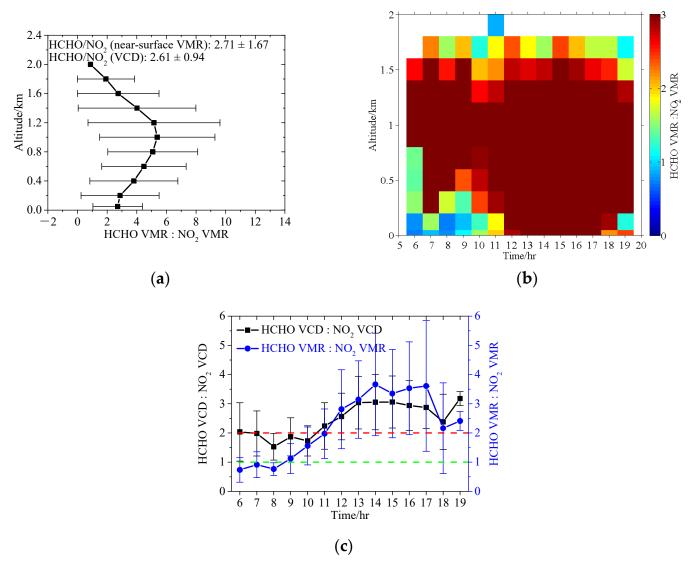


Figure 6. Averaged vertical profile of (a) VMR ratio of HCHO to NO_2 . Diurnal variation of (b) vertical profiles of VMR ratios of HCHO to NO_2 as well as (c) HCHO/ NO_2 ratios of VCD and VMR in the lowest layer. The color blocks are not shown in (b), once the HCHO VMRs are smaller than 1ppb or larger than 10 ppb. The 1:1 and 2:1 ratios of HCHO/ NO_2 in (c) are also shown by the colored dashed lines.

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The averaged diurnal variations of daytime $R_{HCHO-NO2}$ profile are shown in Figure 6b. Smaller $R_{HCHO-NO2}$ (<2) values mainly appear before 11:00 BJ in the lower layer. The height of the lower boundary of the elevated $R_{HCHO-NO2}$ layer gradually decreases from 6:00 BJ (0.8 km) to 11:00 BJ (0.2 km) until the higher $R_{HCHO-NO2}$ (>2) fills the whole lower troposphere (below 1.4 km) during the period of 12:00–17:00 BJ. Above the altitude of 1.6 km, most $R_{HCHO-NO2}$ values are in the range of 1–2. After 19:00 BJ, the smaller $R_{HCHO-NO2}$ in the lower layer seems to occur again. Correspondingly, diurnal variations of $R_{HCHO-NO2,VMR}$ are in the range of 0.74–3.66, greater than that of $R_{HCHO-NO2,VCD}$ (1.53–3.18) (Figure 6c). However, the trends of their diurnal variation are similar with the wide peak around 15:00 BJ, although there are larger differences in the absolute value between $R_{HCHO-NO2,VMR}$ and $R_{HCHO-NO2,VCD}$ before 10:00 BJ and after 19:00 BJ. It is worth noting that both $R_{HCHO-NO2,VMR}$ and $R_{HCHO-NO2,VCD}$ are larger than 2 after 10:00 BJ, when the photochemistry reaction is active. This implies that the sensitivities of O_3 production to NO_x and VOCs are the same in the lower troposphere at Raoyang, which can be further confirmed by $R_{HCHO-NO2}$ vertical distribution (Figure 6a).

3.3.2. Relationship to O₃ Production and Meteorological Condition

The relationships of the daytime (7:00–18:00 BJ) averages of HCHO and NO₂ to O₃ production and meteorological condition are shown in Figure 7. To analyze the influence of HCHO and NO₂ on O₃ production, we assume the daytime surface O₃ diurnal variations at Raoyang station were mainly caused by O_3 photochemical reaction [4,5], using the variation range of daytime O₃ concentration as a proxy for O₃ production ("O₃ Range" in Figure 7a,b). Upon previous studies [39,40], O_3 production regimes could be assigned as VOC-limited $(R_{HCHO-NO2} < 1)$, transition zone $(1 < R_{HCHO-NO2} < 2)$ and NOx-limited $(R_{HCHO-NO2} > 2)$ for HCHO VMRs lower than 10ppb. Therefore, inferred from the R_{HCHO-NO2,VCD} in Figure 7a, O₃ production regimes are in the NOx-limited section and transition zone, corresponding to the larger and smaller O_3 ranges with the typical boundary of ~60 ppb, respectively. Differently from the VCD situation, in which there is no case of R_{HCHO-NO2,VCD} < 1, there is a small quantity of cases with R_{HCHO-NO2,VMR} < 1 (i.e., VOC-limited) as shown in Figure 7b, which is associated with weak near-surface O_3 production (O_3 range < 40 ppb). It is interesting that the strongest near-surface O_3 production (O_3 Range > 90 ppb) occurs in the transition zone, i.e., $1 < R_{HCHO-NO2,VMR} < 2$, probably signifying combined air pollution in the region of Raoyang. On the whole, the cases with high O₃ ranges are found in the NOx-limited regime for both VCDs and near-surface VMRs in summer in the NCP rural area.

In the interests of consistency with surface meteorological observation, we focus on the near-surface HCHO and NO2 VMRs as well as R_{HCHO-NO2,VMR} in the following discussion. It is clear that the zones of $R_{HCHO-NO2,VMR} > 2$ and $R_{HCHO-NO2,VMR} < 1$ are corresponding to the high and low daytime averaged temperatures, respectively (Figure 7c). With the opposite situation for RH, the low (typically 55%) and high (typically 65%) RH appears in the NOx-limited and VOC-limited O₃ production regimes, respectively (Figure 7d). Therefore temperature and RH are two important driving factors of HCHO and NO₂ photochemical process in summer, partly because they affect the photochemistry activity and biogenic isoprene emissions [41,42]. The relationship between R_{HCHO-NO2,VMR} and wind speed is unapparent (Figure 7e), partly because the wind speed changes in a small range. However, the rose of R_{HCHO-NO2,VMR} in 16 wind sectors clearly presents its dependency of wind direction (Figure 7f). The total frequencies for R_{HCHO-NO2,VMR} > 2, $1 < R_{HCHO-NO2,VMR} < 2$, $R_{HCHO-NO2,VMR} < 1$ are 49%, 40% and 11%, respectively. The ratios of R_{HCHO-NO2,VMR} > 2 are distributed at the southeast side of the NE-SW axis with a maximum frequency of \sim 16% in the ESE section. The ratios of $R_{HCHO-NO2,VMR}$ < 1 are mainly distributed in the NNW section with a maximum frequency of ~5%. Therefore, air masses from different source regions can lead to different sensitivities of O₃ production to NO_x and VOCs at the Raoyang station [22]. In addition, southeasterly winds occur mainly during the afternoon period (Figure 5h), when O₃ production is NO_x-limited

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 $(R_{HCHO-NO2}, Figure 6b)$. This can partly explain the dependence of $R_{HCHO-NO2}$ on wind direction (Figure 7f).

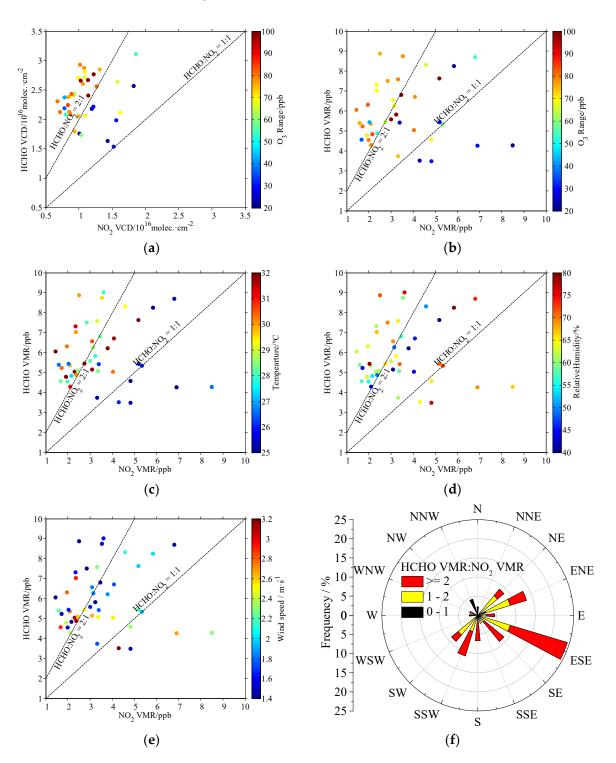


Figure 7. Relationship of HCHO and NO_2 to O_3 and meteorological conditions. (a) HCHO vs. NO_2 VCD daytime (7:00–18:00 BJ) averages, colored by the range of surface O_3 daytime variation. (b) is same as (a), but for HCHO vs. NO_2 VMR daytime averages in the lowest layer. (c-e) is similar to (b), but separately colored by daytime averages of air temperature, relative humidity and wind speed. (f) Rose showing the frequency of VMR ratios of HCHO to NO_2 in 16 wind sectors. The ratios of HCHO/ NO_2 = 1:1 and 2:1 are also presented by the dash lines in (a-e).

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4. Discussion

The level of NO₂ VCD in this study is higher than that at the NCP background station of Shangdianzi and lower than that in the NCP megacity of Beijing in summer [30,32,43]. However, the level of HCHO VCD is higher than that at another NCP rural site of Gucheng and close to the HCHO abundance in Beijing in the summer of 2017 [44]. Satellite observation also found significant spatial differences and temporal variations for NO₂ and HCHO levels over the NCP [45,46]. Even so, we found similar vertical structures and diurnal variations of NO₂ and HCHO profiles retrieved by MAX-DOAS measurements at the NCP rural site of Xingtai [16]. The boundary layer evolution and photochemical reaction played an important role in the diurnal variation of NO₂ and HCHO vertical distribution, leading to the elevated pollution layer in the residual layer before 10:00 BJ and the HCHO oxidation layer in the upper part of the boundary layer from 11:00–17:00 BJ [2,14]. Different from the studies at another NCP suburban site [24], the NO_x -limited regime of O_3 production stayed stable in the boundary layer as a whole. Meteorological conditions also affect the levels and spatiotemporal variation of NO₂ and HCHO, and the sensitivities of O₃ production vary accordingly because meteorological conditions such as temperature can influence the natural source emission and the photochemical efficiency of O₃ precursors [13,23,41].

5. Conclusions

In this article, we present vertical profiles of NO_2 , HCHO and HCHO/ NO_2 ratios retrieved from ground-based MAX-DOAS measurements in summer 2014 at the Raoyang rural site of the NCP. Temporal variations of NO_2 and HCHO vertical profiles as well as the relationship of HCHO/ NO_2 to O_3 production were investigated. The main findings are summarized as follows.

The average levels of VCDs and near-surface VMRs were $1.02 \pm 0.51 \times 10^{16}$ molec·cm $^{-2}$ and 3.23 ± 2.70 ppb for NO₂ and $2.32 \pm 0.56 \times 10^{16}$ molec·cm $^{-2}$ and 5.62 ± 2.11 ppb for HCHO during the field experiment. The NO₂ vertical distribution averaged by all effective profiles had a declining shape with increasing altitude, while there was an elevated layer for HCHO at the altitude of ~1km, indicating an enhancement of photochemistry in the upper planetary boundary layer. Thus, the correlation coefficient between VCDs and near-surface VMRs for NO₂ (R = 0.69) is larger than that for HCHO (R = 0.55), implying more limitations of in situ surface measurement in representation of the vertical variations of HCHO than NO₂ due to HCHO secondary production at higher altitudes. NO₂ and HCHO levels are closely connected with meteorological conditions. For example, higher NO₂ VCDs are usually associated with lower temperature, higher RH and lower PBLH.

The diurnal variations of VCD and near-surface VMR were in the range of 0.74– 1.78×10^{16} molec·cm⁻² and 1.74–7.63 ppb for NO₂ and 2.11– 2.95×10^{16} molec·cm⁻² and 4.67–7.47 ppb for HCHO during the effective observation period, respectively. The severe NO₂ pollution cases, i.e., peaks for VCDs and near-surface VMRs, appeared in the early morning and evening. The higher HCHO VCDs occurred in the morning and evening, while the near-surface HCHO VMRs gradually increased and peaked around 10:00 BJ. There were elevated pollution layers for both NO₂ and HCHO in the residual layer before 10:00 BJ. However, the NO₂ pollution without an elevated layer mainly concentrated at heights less than 0.2 km from 11:00–17:00 BJ, while the elevated HCHO layers always existed, and the associated HCHO concentration levels were lower due to the higher PBLH. Complex effects of boundary layer processes and photochemistry reactions on the diurnal variations of NO₂ and HCHO vertical distribution need to be further studied in the future.

The averaged diurnal variations of daytime HCHO/NO $_2$ ratio ($R_{HCHO-NO2}$) profile indicated that $R_{HCHO-NO2}$ was generally larger than 2 in the boundary layer from 11:00 BJ until 19:00 BJ, when ozone photochemistry was most active. The most frequent O_3 production was in the NOx-limited regime during the observation campaign, inferred from both $R_{HCHO-NO2,VCD}$ and $R_{HCHO-NO2,VMR}$. The ozone-NO $_x$ -HCHO production sensitivities were closely related to the meteorological conditions, such as temperatures, RH and wind. Temperature and RH probably affected photochemistry activity and natural

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source emissions, and the air masses from different source regions transported different atmospheric compositions to Raoyang. Therefore, more investigations of NO_2 and HCHO vertical profiles are still needed to reveal the formation mechanism and long-term evolution of O_3 pollution and to develop strategies for controlling the O_3 photochemical pollution over the NCP in the future.

Supplementary Materials: The following supporting information can be downloaded at: https://www.mdpi.com/article/10.3390/atmos13060860/s1, Figure S1: time series of the qualified hourly NO₂ and HCHO as well as coincident meteorological conditions; Figure S2: time series for vertical profiles of VMR ratios of HCHO to NO₂ as well as HCHO/NO₂ ratios of VCD and near-surface VMR; Table S1: fit settings for the NO₂ and HCHO spectral analyses.

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References

- 1. Seinfeld, J.H.; Pandis, S.N. *Atmospheric Chemistry and Physics: From Air Pollution to Climate Change*, 3rd ed.; John Wiley & Sons, Inc.: Hoboken, NJ, USA, 2016; pp. 175–264.
- 2. Ma, J.Z.; Wang, W.; Chen, Y.; Liu, H.J.; Yan, P.; Ding, G.A.; Wang, M.L.; Sun, J.; Lelieveld, J. The IPAC-NC field campaign: A pollution and oxidization pool in the lower atmosphere over Huabei, China. *Atmos. Chem. Phys.* **2012**, 12, 3883–3908. [CrossRef]
- 3. Ma, J.; Xu, X.; Zhao, C.; Yan, P. A review of atmospheric chemistry research in China: Photochemical smog, haze pollution, and gas-aerosol interactions. *Adv. Atmos. Sci.* **2012**, 29, 1006–1026. [CrossRef]
- 4. Ma, J.; Zhou, X. Summertime tropospheric ozone over China simulated with a regional chemical transport model 2. Source contributions and budget. *J. Geophys. Res.* **2002**, *107*, ACH 2-1–ACH 2-11. [CrossRef]
- 5. Ma, J.; Liu, H.; Hauglustaine, D. Summertime tropospheric ozone over China simulated with a regional chemical transport model 1. Model description and evaluation. *J. Geophys. Res.* **2002**, *107*, 4660. [CrossRef]
- 6. Granier, C.; Bessagnet, B.; Bond, T.; D'Angiola, A.; Denier van der Gon, H.; Frost, G.J.; Heil, A.; Kaiser, J.W.; Kinne, S.; Klimont, Z.; et al. Evolution of anthropogenic and biomass burning emissions of air pollutants at global and regional scales during the 1980–2010 period. *Clim. Change* **2011**, *109*, 163–190. [CrossRef]
- 7. Zhao, B.; Wang, P.; Ma, J.Z.; Zhu, S.; Pozzer, A.; Li, W. A high-resolution emission inventory of primary pollutants for the Huabei region, China. *Atmos. Chem. Phys.* **2012**, *12*, 481–501. [CrossRef]
- 8. Lee, D.S.; Köhler, I.; Grobler, E.; Rohrer, F.; Sausen, R.; Gallardo-Klenner, L.; Olivier, J.G.J.; Dentener, F.J.; Bouwman, A.F. Estimations of global NO_x emissions and their uncertainties. *Atmos. Environ.* **1997**, *31*, 1735–1749. [CrossRef]
- 9. Stavrakou, T.; Müller, J.F.; De Smedt, I.; Van Roozendael, M.; van der Werf, G.R.; Giglio, L.; Guenther, A. Evaluating the performance of pyrogenic and biogenic emission inventories against one decade of space-based formaldehyde columns. *Atmos. Chem. Phys.* **2009**, *9*, 1037–1060. [CrossRef]
- 10. Jin, X.M.; Holloway, T. Spatial and temporal variability of ozone sensitivity over China observed from the Ozone Monitoring Instrument. *J. Geophys. Res. Atmos.* **2015**, 120, 7229–7246. [CrossRef]

Atmosphere 2022, 13, 860 14 of 15

11. Liu, H.; Liu, C.; Xie, Z.; Li, Y.; Huang, X.; Wang, S.; Xu, J.; Xie, P. A paradox for air pollution controlling in China revealed by "APEC Blue" and "Parade Blue". Sci. Rep. 2016, 6, 34408. [CrossRef]

- 12. Sun, Y.; Liu, C.; Palm, M.; Vigouroux, C.; Notholt, J.; Hu, Q.; Jones, N.; Wang, W.; Su, W.; Zhang, W.; et al. Ozone seasonal evolution and photochemical production regime in the polluted troposphere in eastern China derived from high-resolution Fourier transform spectrometry (FTS) observations. *Atmos. Chem. Phys.* **2018**, *18*, 14569–14583. [CrossRef]
- 13. Zhang, X.; Yan, B.; Du, C.; Cheng, C.; Zhao, H. Quantifying the interactive effects of meteorological, socioeconomic, and pollutant factors on summertime ozone pollution in China during the implementation of two important policies. *Atmos. Pollut. Res.* **2021**, 12, 101248. [CrossRef]
- 14. Tang, G.; Liu, Y.; Huang, X.; Wang, Y.; Hu, B.; Zhang, Y.; Song, T.; Li, X.; Wu, S.; Li, Q.; et al. Aggravated ozone pollution in the strong free convection boundary layer. *Sci. Total Environ.* **2021**, *788*, 147740. [CrossRef] [PubMed]
- 15. Liu, C.; Gao, M.; Hu, Q.; Brasseur, G.P.; Carmichael, G.R. Stereoscopic Monitoring: A Promising Strategy to Advance Diagnostic and Prediction of Air Pollution. *B. Am. Meteorol. Soc.* **2021**, *102*, E730–E737. [CrossRef]
- 16. Wang, Y.; Dörner, S.; Böhnke, S.; De Smedt, I.; Dickerson, R.R.; Dong, Z.; He, H.; Li, Z.; Li, Z.; et al. Vertical profiles of NO₂, SO₂, HONO, HCHO, CHOCHO and aerosols derived from MAX-DOAS measurements at a rural site in the central western North China Plain and their relation to emission sources and effects of regional transport. *Atmos. Chem. Phys.* **2019**, *19*, 5417–5449. [CrossRef]
- 17. Liu, R.; Feng, T.; Guo, Y.; Wang, H.; Ming, R.; Wang, S.; Zhou, B. Study on ozone sensitivity in typical tropical region from DOAS ground-to-space measurements. *Acta Sci. Circumstantiae* **2021**, *41*, 2262–2271. (In Chinese)
- 18. Luo, Y.; Dou, K.; Fan, G.; Huang, S.; Si, F.; Zhou, H.; Wang, Y.; Pei, C.; Tang, F.; Yang, D.; et al. Vertical distributions of tropospheric formaldehyde, nitrogen dioxide, ozone and aerosol in southern China by ground-based MAX-DOAS and LIDAR measurements during PRIDE-GBA 2018 campaign. *Atmos. Environ.* **2020**, 226, 117384. [CrossRef]
- 19. Ren, B.; Xie, P.; Xu, J.; Li, A.; Qin, M.; Hu, R.; Zhang, T.; Fan, G.; Tian, X.; Zhu, W.; et al. Vertical characteristics of NO₂ and HCHO, and the ozone formation regimes in Hefei, China. *Sci. Total Environ.* **2022**, *823*, 153425. [CrossRef]
- Tanvir, A.; Javed, Z.; Jian, Z.; Zhang, S.; Bilal, M.; Xue, R.; Wang, S.; Bin, Z. Ground-Based MAX-DOAS Observations of Tropospheric NO₂ and HCHO During COVID-19 Lockdown and Spring Festival Over Shanghai, China. Remote Sens. 2021, 13, 488. [CrossRef]
- 21. Xue, J.; Zhao, T.; Luo, Y.; Miao, C.; Su, P.; Liu, F.; Zhang, G.; Qin, S.; Song, Y.; Bu, N.; et al. Identification of ozone sensitivity for NO₂ and secondary HCHO based on MAX-DOAS measurements in northeast China. *Environ. Int.* **2022**, *160*, 107048. [CrossRef]
- Zong, R.; Yang, X.; Wen, L.; Xu, C.; Zhu, Y.; Chen, T.; Yao, L.; Wang, L.; Zhang, J.; Yang, L.; et al. Strong ozone production at a rural site in the North China Plain: Mixed effects of urban plumesand biogenic emissions. *J. Environ. Sci.* 2018, 71, 261–270. [CrossRef] [PubMed]
- 23. Tan, Z.; Lu, K.; Dong, H.; Hu, M.; Li, X.; Liu, Y.; Lu, S.; Shao, M.; Su, R.; Wang, H.; et al. Explicit diagnosis of the local ozone production rate and the ozone-NO_X-VOC sensitivities. *Sci. Bull.* **2018**, *63*, 1067–1076. [CrossRef]
- 24. Chi, X.; Liu, C.; Xie, Z.; Fan, G.; Wang, Y.; He, P.; Fan, S.; Hong, Q.; Wang, Z.; Yu, X.; et al. Observations of ozone vertical profiles and corresponding precursors in the low troposphere in Beijing, China. *Atmos. Res.* **2018**, 213, 224–235. [CrossRef]
- 25. Cheng, S.; Jin, J.; Ma, J.; Xu, X.; Ran, L.; Ma, Z.; Chen, J.; Guo, J.; Yang, P.; Wang, Y.; et al. Measuring the Vertical Profiles of Aerosol Extinction in the Lower Troposphere by MAX-DOAS at a Rural Site in the North China Plain. *Atmosphere* **2020**, *11*, 1037. [CrossRef]
- 26. Zhang, L.; Yan, P.; Mao, J.; Zhang, X.; Tian, P.; Chang, H. Observation study on aerosol scattering phase function at Raoyang of Hebei, China. *J. Appl. Meteorol. Sci.* **2017**, *28*, 436–446. (In Chinese)
- 27. Wang, R.; Xu, X.; Jia, S.; Ma, R.; Ran, L.; Deng, Z.; Lin, W.; Wang, Y.; Ma, Z. Lower tropospheric distributions of O₃ and aerosol over Raoyang, a rural site in the North China Plain. *Atmos. Chem. Phys.* **2017**, *17*, 3891–3903. [CrossRef]
- 28. Ran, L.; Deng, Z.; Xu, X.; Yan, P.; Lin, W.; Wang, Y.; Tian, P.; Wang, P.; Pan, W.; Lu, D. Vertical profiles of black carbon measured by a micro-aethalometer in summer in the North China Plain. *Atmos. Chem. Phys.* **2016**, *16*, 10441–10454. [CrossRef]
- 29. Peng, W.; Wang, Y.; Gao, X.; Jia, S.; Xu, X.; Cheng, H.; Meng, Z. Characteristics of ambient formaldehyde at two rural sites in the North China Plain in summer. *Res. Environ. Sci.* **2016**, *29*, 1119–1127. (In Chinese)
- 30. Cheng, S.; Ma, J.; Cheng, W.; Yan, P.; Zhou, H.; Zhou, L.; Yang, P. Tropospheric NO₂ vertical column densities retrieved from ground-based MAX-DOAS measurements at Shangdianzi regional atmospheric background station in China. *J. Environ. Sci.* **2019**, *80*, 186–196. [CrossRef]
- 31. Ma, J.; Dörner, S.; Donner, S.; Jin, J.; Cheng, S.; Guo, J.; Zhang, Z.; Wang, J.; Liu, P.; Zhang, G.; et al. MAX-DOAS measurements of NO₂, SO₂, HCHO, and BrO at the Mt. Waliguan WMO GAW global baseline station in the Tibetan Plateau. *Atmos. Chem. Phys.* **2020**, *20*, 6973–6990. [CrossRef]
- 32. Ma, J.Z.; Beirle, S.; Jin, J.L.; Shaiganfar, R.; Yan, P.; Wagner, T. Tropospheric NO₂ vertical column densities over Beijing: Results of the first three years of ground-based MAX-DOAS measurements (2008–2011) and satellite validation. *Atmos. Chem. Phys.* **2013**, 13, 1547–1567. [CrossRef]
- 33. Jin, J.; Ma, J.; Lin, W.; Zhao, H.; Shaiganfar, R.; Beirle, S.; Wagner, T. MAX-DOAS measurements and satellite validation of tropospheric NO₂ and SO₂ vertical column densities at a rural site of North China. *Atmos. Environ.* **2016**, *133*, 12–25. [CrossRef]
- 34. Platt, U.; Stutz, J. Differential Optical Absorption Spectroscopy, Principles and Applications; Springer: Berlin, Germany, 2008; pp. 135–175.

Atmosphere 2022, 13, 860 15 of 15

35. Hönninger, G.; Friedeburg, C.v.; Platt, U. Multi axis differential optical absorption spectroscopy (MAX-DOAS). *Atmos. Chem. Phys.* **2004**, *4*, 231–254. [CrossRef]

- 36. Wang, Y.; Lampel, J.; Xie, P.; Beirle, S.; Li, A.; Wu, D.; Wagner, T. Ground-based MAX-DOAS observations of tropospheric aerosols, NO₂, SO₂ and HCHO in Wuxi, China, from 2011 to 2014. *Atmos. Chem. Phys.* **2017**, 17, 2189–2215. [CrossRef]
- Ryan, R.G.; Rhodes, S.; Tully, M.; Schofield, R. Surface ozone exceedances in Melbourne, Australia are shown to be under NO_x control, as demonstrated using formaldehyde:NO₂ and glyoxal:formaldehyde ratios. Sci. Total Environ. 2020, 749, 141460. [CrossRef]
- 38. Hong, Q.; Liu, C.; Hu, Q.; Zhang, Y.; Xing, C.; Su, W.; Ji, X.; Xiao, S. Evaluating the feasibility of formaldehyde derived from hyperspectral remote sensing as a proxy for volatile organic compounds. *Atmos. Res.* **2021**, 264, 105777. [CrossRef]
- 39. Duncan, B.N.; Yoshida, Y.; Olson, J.R.; Sillman, S.; Martin, R.V.; Lamsal, L.; Hu, Y.; Pickering, K.E.; Retscher, C.; Allen, D.J.; et al. Application of OMI observations to a space-based indicator of NO_x and VOC controls on surface ozone formation. *Atmos. Environ.* **2010**, *44*, 2213–2223. [CrossRef]
- 40. Souri, A.H.; Nowlan, C.R.; Wolfe, G.M.; Lamsal, L.N.; Chan Miller, C.E.; Abad, G.G.; Janz, S.J.; Fried, A.; Blake, D.R.; Weinheimer, A.J.; et al. Revisiting the effectiveness of HCHO/NO₂ ratios for inferring ozone sensitivity to its precursors using high resolution airborne remote sensing observations in a high ozone episode during the KORUS-AQ campaign. *Atmos. Environ.* **2020**, 224, 117341. [CrossRef]
- 41. Duncan, B.N.; Yoshida, Y.; Damon, M.R.; Douglass, A.R.; Witte, J.C. Temperature dependence of factors controlling isoprene emissions. *Geophys. Res. Lett.* **2009**, *36*, L05813. [CrossRef]
- 42. Archibald, A.T.; Neu, J.L.; Elshorbany, Y.F.; Cooper, O.R.; Young, P.J.; Akiyoshi, H.; Cox, R.A.; Coyle, M.; Derwent, R.G.; Deushi, M.; et al. Tropospheric Ozone Assessment Report: A critical review of changes in the tropospheric ozone burden and budget from 1850 to 2100. *Elem. Sci. Anthropocene* 2020, 8, 34. [CrossRef]
- 43. Hendrick, F.; Müller, J.-F.; Clémer, K.; Wang, P.; Mazière, M.D.; Fayt, C.; Gielen, C.; Hermans, C.; Ma, J.Z.; Pinardi, G.; et al. Four years of ground-based MAX-DOAS observations of HONO and NO₂ in the Beijing area. *Atmos. Chem. Phys.* **2014**, *14*, 765–781. [CrossRef]
- 44. Javed, Z.; Liu, C.; Khokhar, M.F.; Tan, W.; Liu, H.R.; Xing, C.Z.; Ji, X.G.; Tanvir, A.; Hong, Q.Q.; Sandhu, O.; et al. Ground-Based MAX-DOAS Observations of CHOCHO and HCHO in Beijing and Baoding, China. *Remote Sens.* 2019, 11, 1524. [CrossRef]
- 45. Zhu, S.; Li, X.; Yu, C.; Wang, H.; Wang, Y.; Miao, J. Spatiotemporal Variations in Satellite-Based Formaldehyde (HCHO) in the Beijing-Tianjin-Hebei Region in China from 2005 to 2015. *Atmosphere* **2018**, *9*, 5. [CrossRef]
- 46. Hou, Y.F.; Wang, L.T.; Zhou, Y.; Wang, S.X.; Liu, W.L.; Zhu, J.F. Analysis of the tropospheric column nitrogen dioxide over China based on satellite observations during 2008–2017. *Atmos. Pollut. Res.* **2019**, *10*, 651–655. [CrossRef]